

1 **Lipophilicity profiling and cell viability assessment of a selected panel of endocrine disruptors**

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14 *Immobilized Artificial Membranes.*

15 **Abstract**

16 Endocrine disruptors are chemicals widely used worldwide by industries in a variety of applications.  
17 Routinely exposure to these chemicals, even if at low doses, can cause damage effects on human  
18 health. In the present study we evaluated toxic effects of nine chemicals, among which phthalates,  
19 using various cell lines to investigate their capability to interfere with cell proliferation and viability.

20 Alongside, we investigated their affinity for phospholipids to assess the possible passage through  
21 biomembranes. Experimentally determined  $\log k_w^{\text{IAM.MG}}$  values ranged from 1.37 to 3.49 whilst  
22 calculated  $\log k_w^{\text{IAM.DD2}}$  spanned from 1.80 to 5.21, supporting the target contaminants to exhibit  
23 lipophilicity to moderate to very high. The achieved results were related to pharmacokinetic and

24 toxicological properties by ADMET predictor™ and EPI Suite™ software. Triclosan and 4-  
25 Nonylphenol were found to be the most toxic against all cell lines screened, showing an IC<sub>50</sub> of 30  
26 μM for triclosan on human keratinocytes and of 50 μM for 4-Nonylphenol on human colorectal  
27 adenocarcinoma cells. Overall, even if the phthalates showed higher IC<sub>50</sub> values (ranging from 170  
28 μM to 280 μM), we can assert that all contaminants herein tested were able to interfere with cell  
29 growth and viability.

30

## 31 **1. Introduction**

32 Contaminants of emerging concern (CECs) are a class of chemicals that are gathering the attention  
33 of the scientific community due to their ubiquity and persistence in the environment, and their  
34 remarkable bio-accumulative and toxic potential. CECs can be included intentionally or  
35 unintentionally in various consumer goods routinely handled by people and identify a broad and  
36 diversified class of compounds including drugs, chemicals contained in personal care products, agents  
37 used for household cleaning and agricultural purposes, additives exploited in industrial processes.  
38 Among CECs, phthalates (PEs), esters of phthalic acid, are chemicals used to aid the dissolution of  
39 other materials (Annamalai and Vasudevan, 2020), since their presence makes plastics more durable,  
40 flexible, and transparent. Bis(2-ethylhexyl) phthalate (DEHP) and Diisononyl phthalate (DINP) are  
41 currently the most widespread PEs, although, over the years, about a dozen of new phthalates have  
42 been synthesized, including Diisodecyl phthalate (DIDP) and Dibutyl phthalate (DBP) (Wang et al.,  
43 2019; Dutta et al., 2020). High molecular weight (MW) PEs, such as DEHP and DINP are employed  
44 in the manufacture of hundreds of plastic products, such as building materials, adhesives, and flexible  
45 vinyl including flooring and wall coverings.

46 Low MW PEs, such DBP, are used in personal care products such as cosmetics, shampoos, skin  
47 moisturizers, hair sprays, nail polish and even in food packaging, instead. Regardless their MW, PEs

48 pose a potential health concern because these can exert a disruptive effect on the endocrine system of  
49 living organisms, due to their action in affecting the hormone homeostasis (Zarean et al., 2016; De  
50 Toni et al., 2017; Zhu et al., 2018; Zhang et al., 2021). Lower MW PEs are rapidly absorbed through  
51 the diet and by inhalation (Başaran et al., 2020; Li et al., 2021), instead PEs with longer alkyl side  
52 chains exhibit low dermal absorption, according to several toxicokinetic studies (Fromme, 2011;  
53 Giuliani et al., 2020). Many scientific reports support that PEs can exert adverse effects on liver,  
54 kidneys, and the reproductive system (Kim et al., 2018; Karwacka et al., 2019; Wang and Qian,  
55 2021).

56 The aim of our research was the investigation of toxicity of five PEs widely used by industries along  
57 with other endocrine disrupting chemicals (EDCs), such as Triclosan, (TCS), 1,2,4,5-  
58 Tetrachlorobenzene (TCB), 1,4-Dichlorobenzene (DCB), 4-Nonylphenol (4-NP), belonging to other  
59 chemical classes with lipophilicity similar or lower than that of the phthalate group. The chemical  
60 structures of the target compounds are showed in Figure 1. To gain access to the receptor site, and,  
61 therefore, trigger harmful effects in the human body, EDCs must necessarily cross the biological  
62 barriers. For this reason, we explored the affinity for membrane phospholipids of all nine chemicals  
63 by Immobilized Artificial Membrane (IAM) Liquid Chromatography (LC) technique (Barbato et al.,  
64 2004)(Grumetto and Russo, 2021), yielding  $\log k_w^{\text{IAM}}$  values, *i.e.* a phospholipophilicity measure.  
65 This method allows to depict distinct interactions, mainly of electrostatic nature, that are different  
66 from those encoded by the “classical” lipophilicity. It represents a crucial parameter for describing  
67 passive transcellular diffusion through biomembranes, that is traditionally expressed by the logarithm  
68 of the ratio of analyte concentrations in an organic solvent, usually *n*-octanol, and an aqueous phase  
69 ( $\log P$ ) (Liu et al., 2011). Moreover, we used the software ADMET Descriptor<sup>TM</sup> (Ghosh et al., 2016)  
70 and Estimation Program Interface (EPI Suite<sup>TM</sup>), to explore possible relationships between  
71 phospholipophilicity and pharmacokinetics parameters as well as ecotoxicological endpoints of all  
72 nine considered molecules. Indeed, recently scientific evidence supported that exposure to CECs such

73 as EDCs could contribute to the pathogenesis of metabolic diseases. In particular, they could cause  
74 developmental and reproductive toxicity, hepatotoxicity, neurotoxicity, and even carcinogenesis (Pan  
75 et al., 2006; Meng et al., 2014). In order to evaluate cellular toxicity potential of EDCs under  
76 investigation in this research work, we carried out *in vitro* experiments in terms of interference with  
77 cell growth and proliferation on a restricted panel of well-established human cells, including both  
78 healthy and cancer cell lines.

79

## 80 **2. Materials and Methods**

### 81 *2.1 Chemicals*

82 Acetonitrile (minimum purity  $\geq 95.0\%$ ) was purchased from Sigma Aldrich (Milan, Italy). Bis(2-  
83 ethylhexyl) phthalate (DEHP- minimum purity  $\geq 98.0\%$ ), Diisononyl phthalate (DINP- minimum  
84 purity  $\geq 99.0\%$ ), Diisodecyl phthalate (DIDP- minimum purity  $\geq 99.0\%$ ), Dibutyl phthalate (DBP-  
85 minimum purity  $\geq 99.0\%$ ), Dioctyl phthalate (DnOP- minimum purity  $\geq 98.0\%$ ), 4-nonylphenol (4-  
86 NP- minimum purity  $\geq 98.0\%$ ), 1,4-dichlorobenzene (DCB- minimum purity  $\geq 99.0\%$ ), 1,2,4,5-  
87 tetrachlorobenzene (TCB- minimum purity  $\geq 98.0\%$ ), Triclosan (TCS- minimum purity  $\geq 97.0\%$ ) were  
88 purchased from Sigma-Aldrich (Dorset, United Kingdom).

### 89 *2.2 IAM chromatography*

90 For DCB, TCB and TCS, phospholipophilicity, measured as  $\log k_w^{IAM}$ , was experimentally  
91 determined by high-performance liquid chromatography (LC-20 AD VP; Shimadzu Corp., Kyoto,  
92 Japan), equipped with an ultraviolet (UV)–visible detector (Shimadzu Model SPD10 AV) set at  $\lambda$  220  
93 nm, analytical column IAM.PC.MG ( $4.6 \times 150$  mm; Regis Chemical Company, Morton Grove, IL).  
94 Mobile phases were vacuum filtered through  $0.45 \mu\text{m}$  nylon membranes (Millipore, Burlington, MA  
95 USA). Data acquisition and integration were accomplished by Cromatoplus software for personal

96 computer. Employed eluents were 0.1 M phosphate buffer at pH 7.0 and acetonitrile at various  
97 percentages with a flow rate ranging from 1.0 to 2.3 mL min<sup>-1</sup>. All samples were dissolved in  
98 acetonitrile (ca. 10<sup>-4</sup> M) and chromatographic analysis were carried out at 22 ± 2° C. Affinity of the  
99 chemicals for the IAM.PC.MG was measured as retention factor extrapolated at 100 % of aqueous  
100 phase ( $k_w^{IAM}$ ) by performing a polycratic method of extrapolation (Braumann et al., 1983). Since all  
101 the compounds under our investigation required at least the addition of acetonitrile to mobile phase  
102 to elute within 20 min, three different mobile phases containing acetonitrile in percentages ( $\phi$ )  
103 ranging from 10% to 30% (v/v) were employed. All values of  $\log k_w^{IAM}$  are the average of at least  
104 three measurements.

### 105 2.3 IAM calculation

106 Six of the studied analytes were not experimentally accessible on the IAM.MG stationary phase,  
107 while all compounds were again not experimentally accessible on the IAM.DD2 phase. For this  
108 reason, the remaining analytical values were predicted *in silico* via the correlative equations reported  
109 in (Russo et al., 2017). The tool (available at <https://www.ddl.unimi.it/vegaol/logkwiam.htm>) offers  
110 a calculation of  $k_w^{IAM}$  values on both IAM.PC.MG and IAM.PC.DD2 columns, of any molecule  
111 included in the PubChem collection as implemented in the script version. The software is based on  
112 205  $\log k_w^{IAM.MG}$  and 161  $\log k_w^{IAM.DD2}$  values experimentally achieved by LC.

### 113 2.4 *In silico* prediction of ADMET properties and ecotoxicological endpoints

114 As the nine chemicals under our investigation differ in physicochemical, as well as absorption,  
115 distribution, metabolism, excretion, and toxicity (ADMET) profile, we used ADMET Predictor™  
116 software for Windows-based personal computers, (version 8.1.0.11, Simulations Plus, Lancaster, CA,  
117 USA), as tool applicable for the estimation of ADMET properties. ECOSAR module of EPI Suite™  
118 Software (Version 4.11., 2012. U.S. Environmental Protection Agency, Office of Pollution,  
119 Prevention and Toxics: Washington) was employed to calculate ecotoxicological endpoints as pLC50

120 values (96-h) for fish. EPI environmental software is a predictive tool, freely available at E.P.A. (U.S.  
121 EPA, 2012), estimating aquatic toxicity-acute (short-term) and chronic (long-term) of chemicals in  
122 aquatic organisms, among which fishes.

### 123 *2.5 Cell cultures*

124 Human breast adenocarcinoma cells MCF-7 (Endocrine-Responsive, ER), human ovarian cancer  
125 cells A2780 and human keratinocytes HaCaT were grown in DMEM (Invitrogen, Paisley, UK)  
126 supplemented with 10% fetal bovine serum (FBS, Cambrex, Verviers, Belgium), L-glutamine (2 mM,  
127 Sigma, Milan, Italy), penicillin (100 units/ml, Sigma) and streptomycin (100 µg/ml, Sigma), and  
128 cultured in a humidified 5% carbon dioxide atmosphere at 37 °C. Human colon adenocarcinoma cells  
129 Caco-2 were grown in Eagle's Minimum Essential Medium (EMEM) supplemented with 20% fetal  
130 bovine serum (FBS, Cambrex, Verviers, Belgium), and cultured in a humidified 5% carbon dioxide  
131 atmosphere at 37 °C. Moreover, Caco-2 cells, led to differentiation, were used at post-confluence  
132 stage as a model of human enterocytes (Di Cesare Mannelli et al., 2018). All cell lines were grown  
133 according to ATCC recommendations.

### 134 *2.6 Treatments and cell toxicity assessment*

135 The cytotoxic activity of selected contaminants was investigated through the estimation of a “cell  
136 survival index”, arising from the combination of cell viability evaluation with cell counting (Russo  
137 et al., 2021). Cells were inoculated in 96-microwell culture plates at a density of  $10^4$  cells/well and  
138 allowed to grow for 24 h. The medium was replaced with fresh medium and cells were treated for  
139 further 48 h with a range of concentrations (10 → 200 µM) of each chemical under our study. Cell  
140 viability was evaluated using the MTT assay, which measures the level of mitochondrial  
141 dehydrogenase activity using the yellow 3-(4,5-dimethyl-2-thiazolyl)-2,5-diphenyl-2H-tetrazolium  
142 bromide (MTT, Sigma) as substrate. The assay is based on the redox ability of living mitochondria  
143 to convert dissolved MTT into insoluble purple formazan. Briefly, after the treatments, the medium

144 was removed, and the cells were incubated with 20  $\mu\text{L}$ /well of a MTT solution (5 mg/mL) for 1 h in  
145 a humidified 5 %  $\text{CO}_2$  incubator at 37 °C. The incubation was stopped by removing the MTT solution  
146 and by adding 100  $\mu\text{L}$ /well of DMSO to solubilize the obtained formazan. Finally, the absorbance  
147 was monitored at 550 nm using a microplate reader (iMark microplate reader, Bio-Rad, Milan, Italy).  
148 Cell number was determined by TC20 automated cell counter (Bio-Rad, Milan, Italy), providing an  
149 accurate and reproducible total count of cells and a live/dead ratio in one step by trypan blue exclusion  
150 assay. Bio-Rad's TC20 automated cell counter uses disposable slides, TC20 trypan blue dye (0.4 %  
151 trypan blue dye w/v in 0.81 % sodium chloride and 0.06 % potassium phosphate dibasic solution)  
152 and a CCD camera to count cells based on the analyses of captured images. Once the loaded slide is  
153 inserted into the slide port, the TC20 automatically focuses on the cells, detects the presence of the  
154 trypan blue dye, and provides the count. When cells are damaged or dead, trypan blue can enter the  
155 cell allowing living cells to be counted. After chemical exposure in 96-microwell culture plates, the  
156 medium was removed, and the cells were collected. 10  $\mu\text{L}$  of cell suspension, mixed with 0.4 % trypan  
157 blue solution at 1:1 ratio, were loaded into the chambers of disposable slides. The results are expressed  
158 in terms of total cell count (number of cells per mL). If trypan blue is detected, the instrument also  
159 accounts for the dilution and shows live cell count and percent viability. Total counts and live/dead  
160 ratio from random samples for each cell line were subjected to comparisons with manual  
161 hemocytometers in control experiments. The calculation of the concentration required to inhibit the  
162 net increase in the cell number and viability by 50% (IC50) is based on plots of data ( $n = 6$  for each  
163 experiment) and repeated five times (total  $n = 30$ ). IC50 values were obtained by means of a  
164 concentration response curve by nonlinear regression using a curve fitting program, GraphPad Prism  
165 8.0, and are expressed as mean values  $\pm$  SEM ( $n = 30$ ) of five independent experiments.

## 166 *2.7 Statistical analysis*

167 All data were presented as mean values  $\pm$  SEM. The statistical analysis was performed using Graph-  
168 Pad Prism (version 8.0, Graph-Pad software Inc., San Diego, CA) and ANOVA test for multiple  
169 comparisons was performed followed by Bonferroni's test.

### 170 3. Results and Discussion

171 Keeping in mind that several environmental contaminants have been associated with biological  
172 processes leading to adverse human health consequences, the nine chemicals *i.e.*, 5 phthalates, 2  
173 chlorobenzenes, 4-Nonylphenol and Triclosan were tested for their ability to interfere with cell  
174 viability. In particular, their *in vitro* biological effects on a selected panel of healthy and tumor cell  
175 lines (HaCaT, enterocytes, MCF-7, A2780 and Caco-2) were investigated, as reported in Figures 2  
176 and 3 and Table 1, the latter showing IC<sub>50</sub> values after 48 h of incubation *in vitro*.

177 Cell toxicity data of the panel of contaminants under investigation were consistent in indicating that  
178 TCS and 4-NP are the most toxic compounds inducing strong effects in all the *in vitro* models used,  
179 showing IC<sub>50</sub> values in the low micromolar range. TCS has shown toxic effects on HaCaT  
180 keratinocytes (IC<sub>50</sub> of 30  $\mu$ M), on MCF-7 breast cancer cells (IC<sub>50</sub> of about 45  $\mu$ M) and on Caco-2  
181 cells (IC<sub>50</sub> of about 50  $\mu$ M), whereas on enterocytes and Caco-2 cells IC<sub>50</sub> values were at 50  $\mu$ M. 4-  
182 NP has shown toxic effects starting from 50  $\mu$ M on all cell lines tested. Previous studies  
183 demonstrated that TCS-induced toxicity is related to its capacity to dysregulate the redox status of  
184 cells, leading to oxidative damage of lipids, proteins, and DNA (Binelli et al., 2009; Park et al., 2016;  
185 Zhang et al., 2018). Moderate toxicity was found for TCB on MCF-7 breast cancer cells and for DCB  
186 on HaCaT cells. Among the phthalates, the toxic effects were moderate. DBP and DEHP moderately  
187 interfered with MCF-7 cells viability. Moreover, we have also observed a contaminants-induced  
188 proliferative effect. In particular, DNOP and DIDP induced a significant increase in A2780 cells  
189 proliferation after 48h of treatment at 50  $\mu$ M, whereas a significant increase of proliferation of  
190 enterocytes was observed after 48h of DEHP exposure at 50  $\mu$ M. Overall, our results agree with the

191 literature (Lepretti et al., 2015; Rubin and Zucker, 2022) that assessed TCS and 4-NP cause a decrease  
192 in cell viability of Caco-2 cells. On MCF-7 cells our results for 4-NP treatment agree with Vivacqua  
193 et al. (Vivacqua et al., 2003), that demonstrated an important inhibition of cell proliferation at 100  
194  $\mu\text{M}$ , although at low concentrations 4-NP induced cell proliferation. Similarly, Yoon et al. (Yoon and  
195 Kwack, 2021) reported a proliferative effect of TCS on MCF-7 cells, suggesting that the  
196 contaminants-induced proliferative effects could be associated with estrogenic activity interference.

197 The results regarding phospholipophilicity values of the nine chemicals under analysis, were  
198 expressed as  $\log k_w^{\text{IAM.MG}}$  and  $\log k_w^{\text{IAM.DD2}}$  values either calculated *in silico* or experimentally  
199 measured on IAM.PC.MG stationary phase. Table 2 shows all data, as well as log P values.

200 IAM.LC approach, both as experimental approach and *in silico* predicted values, was chosen as the  
201 log P determination by the traditional “shake-flask” method was poorly reproducible, due to the very  
202 high lipophilicity values, up to 10.36 of DIDP. Actually, IAM.LC can be considered as a technique  
203 for the determination of the affinity xenobiotics/biomembranes and therefore of their potential of  
204 crossing of biological barriers. Indeed, the IAM stationary phases consist of analogues of  
205 phosphatidylcholine (PC), one of the main constituents of biological membranes. Unfortunately,  
206 despite we tried shorter column formats (as short as 1 cm), it has not been possible to carry out the  
207 determination also on IAM.PC.DD2 stationary phase, due to the very high retention time of the  
208 analytes, which were unsuitable as (phospho) lipophilicity indexes (Taillardat-Bertschinger et al.,  
209 2003).

210 IAM.PC.DD2 differs from IAM.PC.MG in the end-capping of residual amino groups of the silica-  
211 propylamine core supporting  $\text{C}_{10}$  and  $\text{C}_3$  alkyl chains being end-capped by both decanoic and  
212 propionic anhydrides, while IAM.PC.MG supports hydroxy groups, being end-capped by methyl  
213 glycolate. The different chemistry of the end-capping makes the stationary phase IAM.MG more  
214 polar than IAM.DD2, and better suited for the profiling of these contaminants. Anyway, in all the

215 other works by our group, IAM parameters measured on both stationary phases, were found to be  
216 strongly interrelated (Grumetto et al., 2012; Grumetto et al., 2016; Russo et al., 2017). This is accurate  
217 even so more with the compounds belonging to this panel, as all of them are essentially uncharged at  
218 the experimental conditions (pH 7.0). Indeed,  $\log k_w^{\text{IAM}}$  values for structurally unrelated neutral  
219 compounds, was found to unambiguously relate with the respective *n*-octanol/water partition  
220 coefficients ( $\log P$ ), as no electrostatic interactions, which are distinctive of the IAM phase as  
221 compared to *n*-octanol/water, can occur between the analytes and the electrically charged stationary  
222 phase (Grumetto et al., 2016). The  $\log k_w^{\text{IAM}}$  values of DCB, TCB and TCS was achieved  
223 experimentally on IAM.PC.MG, while the  $\log k_w^{\text{IAM}}$  values of the remaining chemicals were  
224 calculated, as inaccessible even at maximum percentage of organic solvent, *i.e.* 30% (v/v) allowed.  
225 IAM indexes ( $\log k_w^{\text{DD}2}$ ) were related to all ADMET Predictor™ parameters. This software includes  
226 several built models based on a statistical methodology that generates, starting from molecular  
227 structure, physicochemical properties with a high degree of accuracy. Designed using artificial neural  
228 network ensemble (ANNE) models (Paixao et al., 2014), it generates 140 *in silico* calculated  
229 properties and predicts toxicological endpoints. Among ADMET™ output, we selected those with  
230 the best scores; four parameters were found quite well related, suggesting a link between  
231 phospholipophilicity and the modulation of the passage through cell membranes, that is a prerequisite  
232 for causing toxicity: a) ADMET Risk, indicating the number of potential ADMET problems a  
233 compound might have (score range 0-22), b) Absorption Risk, indicating the number of potential oral  
234 absorption problems a compound could have (score range 0-8), c) LD50 for rat acute toxicity (mg/Kg  
235 in oral dose that would be lethal to 50% of the rats), d) DiffCoeff, an Hayduk-Laudie infinite dilution  
236 diffusion coefficient ( $\text{cm}^2/\text{s} \cdot 10^5$ ) of nonelectrolytes in water. The best relationships are shown in  
237 Figure 4, while the full matrix is reported in supplementary material section (Table S1). PEs show  
238 lowest diffusion values, except for DBP, which behaves as an outlier, probably due to its lower  
239 phospholipophilicity values and its shorter alkyl chains. IAM.LC measurements relate quite well

240 with permeability and toxicity data calculated *in silico* and offer information that was complementary  
241 to that afforded by cell viability assays.

242 ECOSAR™, included in EPI Suite™, estimates the acute (short-term) and chronic (long-term or  
243 delayed) toxicity of industrial chemicals on aquatic organisms, such as fishes, aquatic invertebrates,  
244 and green algae. Predicted pLC<sub>50</sub> values reported in Table 3 show a good relationship with log kw<sup>IAM</sup>  
245 (r=0.9729), indicating that a higher phospholipophilicity is related to a higher value of toxicity for  
246 fishes (Figure 5).

#### 247 **4. Conclusions**

248 *In vitro* data support a stronger cytotoxic action of two compounds *i.e.*, 4-NP and TCS, on all  
249 investigated cell lines, both normal and cancerous. This is consistent with similar studies by  
250 independent scholar and could be attributed to a perturbation of the redox status of cells, suggesting  
251 a universal rather than a tissue specific mechanism of action. The reduced cell viability induced by  
252 4-NP and TCS can be attributed to optimal phospholipophilicity and diffusion values on the contrary  
253 of PEs that did not show a significant cytotoxicity. The phospholipophilicity  
254 measurements/calculations allowed to achieve a consistent scale of affinity of these compounds for  
255 the cellular membranes. These were found to be related to data of ADMET risk and aquatic toxicity,  
256 as predicted by specialized software.

#### 257 **Declaration of interest**

258 The authors do not declare any competing financial interest.

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