

1 **Title**

2 Thermal sensitivity of the crab *Neosarmatium africanum* in tropical and temperate mangroves on the east coast of Africa.

3

4 **Authors**

5 Marco Fusi^{1,*}, Simone Babbini², Folco Giomi³, Sara Fratini², Farid Dahdouh-Guebas^{4,5}, Daniele Daffonchio¹, Christopher
6 McQuaid⁶, Francesca Porri^{6,7}, Stefano Cannicci^{8,2}

7

8 **Authors' affiliation**

9 ¹King Abdullah University of Science and Technology (KAUST), Biological and Environmental Sciences & Engineering
10 Division (BESE), Thuwal 23955-6900, Saudi Arabia;

11 ²Department of Biology, University of Florence, Via Madonna del Piano 6, 50019 Sesto Fiorentino, Italy;

12 ³Department of Agronomy Food Natural resources Animals and Environment (DAFNAE), University of Padova, Viale
13 dell'Università, 16, 35020 Legnaro, PD, Italy;

14 ⁴Laboratory of Systems Ecology and Resource Management, Department of Organism Biology, Faculty of Sciences,
15 Université Libre de Bruxelles – ULB, Avenue F.D. Roosevelt 50, CPI 264/1, B-1050 Brussels, Belgium.

16 ⁵Ecology & Biodiversity, Laboratorium voor Algemene Plantkunde en Natuurbeheer (APNA), Department of Biology,
17 Faculty of Sciences and Bio-engineering Sciences, Vrije Universiteit Brussel – VUB, Pleinlaan 2, B-1050 Brussels,
18 Belgium

19 ⁶Coastal Research Group, Department of Zoology and Entomology, Rhodes University, Grahamstown, South Africa;

20 ⁷South African Institute for Aquatic Biodiversity (SAIAB), Somerset Street, Grahamstown 6139, South Africa

21 ⁸The Swire Institute of Marine Science and The School of Biological Sciences, The University of Hong Kong, Pokfulam
22 Road, Hong Kong

23 *Corresponding author: marco.fusi@kaust.edu.sa

24

25

26

27

28

29

30 **Abstract**

31 Mangrove forests are among the tropical marine ecosystems most severely affected by rapid environmental change and
32 the activities of key associated macrobenthic species contribute to their ecological resilience. Along the east coast of
33 Africa, the amphibious sesarmid crab *Neosarmatium africanum* (= *meinerti*) plays a pivotal role in mangrove ecosystem
34 functioning through carbon cycling and sediment bioturbation. In the face of rapid climate change, identifying the
35 sensitivity and vulnerability to global warming of this species is of increasing importance. Based on a latitudinal
36 comparison, we measured the thermal sensitivity of an tropical and a temperate population of *N. africanum*, testing
37 specimens at the centre and southern limit of its distribution, respectively. We measured metabolic oxygen consumption
38 and haemolymph dissolved oxygen content during air and water breathing within a temperature range that matched the
39 natural environmental conditions. The results indicate different thermal sensitivities in the physiological responses of *N.*
40 *africanum* from tropical and temperate populations, especially during air breathing. The differences observed in the
41 thermal physiology between the two populations suggest that the effect of global warming on this important mangrove
42 species may be different under different climate regimes.

43

44 **Keywords**

45 Sesarmidae; Decapods; Tropical and Temperate wetlands; Oxygen Consumption; Haemolymph; Physiology, Populations.

46

47 **Introduction**

48 Recent integrative frameworks propose that the vulnerability of species to environmental changes depends on their
49 capacity to individually adapt their physiology and behaviour in response to the changes they effectively experience
50 (Williams et al., 2008; Huey et al., 2012; Rezende et al., 2014). These frameworks thus suggest that the vulnerability of
51 a species to climate change should be assessed through a mechanistic approach, based on the integration of data from its
52 biological traits (such as behaviour, thermal physiology and metabolism) with environmental data (Gaston et al., 2009;
53 Kearney & Porter, 2009; Sih et al., 2010; Kearney et al., 2012). Among physiological traits, respiration and respiration
54 control provide one of the most accurate proxies of a species' thermal sensitivity, defined as the physiological response
55 (such as metabolic oxygen consumption and haemolymph oxygen content) to changes in its thermal environment (Sinclair
56 et al., 2016; Verberk et al., 2016). Furthermore, it is becoming evident that the assessment of thermal sensitivity has to be
57 determined throughout a species entire distributional range, since conspecific populations subject to different
58 environmental conditions can respond in different ways (Eliason et al., 2011; Sunday et al., 2011; Baldanzi et al., 2015;
59 Fusi et al., 2015).

60 Mangrove forests are amongst the most vulnerable and endangered ecosystems in the world (Duke et al., 2007; Hoegh-
61 Guldberg & Bruno, 2010) and are heavily threatened by sea-level rise as a consequence of global warming (Gilman et al.,
62 2008; Lovelock et al., 2015). Their exposure to factors related to climate change, such as rising temperature, is, however,
63 still debated amongst ecologists (Gilman et al., 2008) and few data are available on the relevant ecological traits of key
64 benthic species that play a critical role in mangrove ecosystem functioning (Lee, 2008). In east African mangrove forests,
65 the large burrowing sesarmid crab *Neosarmatium africanum* (namely=*meinerti*, Ragionieri, Fratini & Schubart, 2012),
66 can occur at densities of over 20 individuals per square meter (Andreetta et al., 2014). *N. africanum* is a semi-terrestrial
67 species and is the African representative of a complex of four sister species colonizing the Indian Ocean and East Australia
68 regions (Ragionieri et al., 2010, 2012). It occupies the landward fringe of mangrove forests (Macnae, 1968; Hartnoll,
69 1975), which is inundated only during spring tides and is frequently dominated by *Avicennia marina* trees (Forssk.) Vierh.
70 (Tomlinson, 1986). *N. africanum* provides crucial ecological functions for the entire ecosystem such as burrowing
71 (Micheli et al., 1991; Berti et al., 2008) and a contribution to carbon burial and storage (Andreetta et al., 2014), among
72 the others. By burrowing, *N. africanum* also contributes to modification of sediment topography and the distribution of
73 sediment grain size (Warren & Underwood, 1986), reduces pore water salinity (Ridd, 1996; Stieglitz et al., 2000), creates
74 microhabitats for other fauna (Bright & Hogue, 1972; Dittmann, 1996; Tack et al., 2001), contributes to secondary
75 production (Lee, 1997) and increases nutrient levels while decreasing sulfide concentrations in the sediment (Smith et al.,
76 1991; Kristensen, 2008).

77 Despite such an array of multiple and critical functions, information on the sensitivity of *N. africanum* to environmental
78 changes is lacking. The landward edge of the *A. marina* zone is a particularly harsh environment for intertidal species as
79 it is subject to acute fluctuations in both salinity and temperature (Macnae & Kalk, 1962). Gillikin et al. (2004) showed
80 that these crabs are highly effective hyper/hypo-osmoregulators, able to survive a range of 16–65 ppt of salinity, but no
81 data are available concerning this species' thermal responses. Given the accumulating evidence that allopatric conspecific
82 populations may exhibit important differences in their metabolic responses to stress, we asked whether *N. africanum*
83 individuals belonging to tropical and temperate populations may have a different thermal sensitivity. We addressed this
84 question by examining physiological responses of *N. africanum* to temperature in the laboratory and coupling this result
85 with the temperatures they experience in the field. Since a population comparison is fundamental to a reliable assessment
86 of species sensitivity (Eliason et al., 2011; Fusi et al., 2015), we carried out these experiments on specimens at the center
87 (Kenya) and the southern limit (South Africa) of *N. africanum* distribution.

88

89 **Materials and methods**

90

91 *Study areas (Fig. 1)*

92 The study was performed during the hottest season at each study site, November/December in Kenya as the tropical site,
93 January/February in South Africa as the temperate site.

94

95 Tropical site: Kenya – Gazi Bay (4° 22' S, 39° 30' E)

96 Gazi bay is a semi-enclosed shallow coastal system located about 40 km south of Mombasa, Kenya. The climate is
97 typically monsoonal, with moist southeast monsoons from March to September and dry northeast monsoons from October
98 to February; rain occurs from March to May and, to a lesser extent, during October and November (Kitheka et al., 1996)
99 Average annual maximum temperature value is around 27-30°C throughout the year.

100

101 Temperate site: South Africa – Mngazana estuary (31°42' S, 29°25' E)

102 The Mngazana River is situated about 250km south-west of Durban, on the southeast coast of South Africa. The estuary
103 measures 5.3km in length and is permanently open to the sea. There are two creeks, which support the main populations
104 of mangroves. Rainfall occurs throughout the year but especially during summer (November– January) (Rajkaran &
105 Adams, 2012) Temperatures vary from an average maximum values of 30-33°C in summer to 10-14°C in winter. The
106 mangrove forest at the Mngazana estuary is one of the southerly in the world (Quisthoudt et al., 2013).

107

108 *Thermal and tidal series*

109 To determine the temperature range experienced by natural populations of *N. africanum*, temperature was recorded in the
110 field for approximately two months during the summer period (49 days in Kenya, 43 days in South Africa) in 2011/2012.
111 Forty temperature loggers (Maxim integrated product, ColdChain Thermodynamics) were placed in areas inhabited by
112 the focal species, avoiding direct exposure to the sun; 20 were positioned about 3 cm above the sediment surface and 20
113 approximately 20 cm beneath the sediment surface close to the burrows of the animals in order to record the temperature
114 to which the animals were directly exposed (Edney, 1961). Loggers were waterproofed using silicon cases that do not
115 affected the accuracy of measurements (Roznik & Alford, 2012). Four additional loggers were used to measure air
116 temperature and humidity. These were attached to branches under the canopy, about 4 m from the ground to avoid tidal
117 submersion, and were protected from rain with plastic covers. All loggers were set up to measure temperature ($\pm 0.1^\circ\text{C}$)
118 at five minute intervals for 15 days, after which data were downloaded and the loggers reset and re-deployed for a total
119 of 4 times in Kenya, 3 in South Africa. Data were downloaded with Cold Chain Thermodynamics software (version 4.9
120 - Fairbridge Technologies) and average temperature was calculated for every hour of all recorded days.
121 Tidal series data were retrieved by Wtide software version 3.1.7 (www.wtide.com) taking as reference points Kilindini
122 and East London in Kenya and South Africa, respectively, corrected with the delay recorded for our study sites. The tides
123 were therefore correlated hourly with the temperature.

124

125 *Crab sampling for laboratory experiments*

126 Sixteen adult male *N. africanum* of similar size (approximately 40mm carapace width) were collected at each site. For
127 acclimation, they were held for two days in aquaria prior to the start of the experiments. They were kept in filtered
128 seawater (35‰ salinity) at $27 \pm 0.5^\circ\text{C}$, under a 12h/12h light/dark cycle. In Kenya, the animals were kept at the laboratory
129 of KMFRI (Kenya Marine and Fisheries Research Institute) in Gazi, while in South Africa at the Coastal Research Group
130 Laboratory, Rhodes University, Grahamstown.

131

132 *Oxygen consumption*

133 Oxygen consumption (MO_2) approximating the routine metabolic rate, was measured in air and water for 8 adult males
134 for each site using an intermittent flow respirometer equipped with eight parallel darkened Perspex chambers placed in a
135 temperature controlled water bath. An oxygen sensor (Sensor Type PSt3 PreSens, Regensburg, Germany), glued to the
136 inside wall of the chamber and connected to a single channel oxygen transmitter Fibox 3 (PreSens, Regensburg, Germany)
137 through an optical sub miniature fiber, was used to measure the partial pressure of oxygen in air and water. Data were
138 recorded using the FibSoft v.1.0 software (Loligo Systems ApS). Prior to measurements, sensors were calibrated in air-

139 equilibrated seawater (100% oxygen saturation) and in sodium thiocyanate saturated seawater (0% oxygen). During trials,
140 oxygen concentration was not allowed to fall below 60% in order to avoid exposing the animals to severe hypoxic
141 conditions (Schurmann & Steffensen, 1992). The limited movements of individuals inside the experimental chambers
142 were adequate to ensure stirring of the water and MO_2 was determined by measuring the linear decline in oxygen
143 saturation. An empty chamber was used during each trial as a control, to account for background oxygen depletion, which
144 was less than 2% of the animals' consumption in water and negligible in air. Prior to ramping of temperature, individuals
145 were placed in the chambers and allowed to recover from handling stress overnight at $27 \pm 0.5^\circ\text{C}$. From an initial
146 measurement performed at 27°C , MO_2 was determined at every two degrees of temperature across the increasing range
147 27 to 37°C , raising temperature at the rate of 1°C h^{-1} (Terblanche et al., 2011). Differences in the variability of MO_2 were
148 not caused by differences in behaviour between the two populations (personal observation). Following each experiment,
149 every animal was individually weighed and its volume calculated by immersing it in a graduated cylinder and measuring
150 the water displacement. All experiments lasted less than 24 h to avoid interference with the metabolic rate by other factors
151 such as starvation and during air respiration humidity was kept at 90% to avoid desiccation (Terblanche et al., 2011).
152 Since the Q_{10} coefficient is an integrated measure of biochemical reactions with physical processes in relation to increases
153 in temperature, it provides a good proxy for thermal performance. $Q_{10[27-37^\circ\text{C}]}$ were therefore calculated for each treatment
154 following Baldanzi et al. (2015).

155

156 *Haemolymph dissolved oxygen content*

157 Dissolved oxygen content in arterial and venous (*sensu* Greenaway & Farrelly, 1984) haemolymph was measured with
158 fiber-optic oxygen microsensors (PreSens GmbH) connected to an oxygen meter (MicrOx-TX, PreSens GmbH) with
159 integrated signal processing software. Sensors were calibrated before each experiment using a two-point calibration in
160 oxygen-free (addition of sodium dithionite) and air-saturated seawater. Animals were acclimated overnight at 27°C and
161 the oxygen dissolved in haemolymph was estimated between 27°C and 38°C (accuracy $\pm 1^\circ\text{C}$), using the protocol
162 described for the MO_2 experiment.

163 Arterial haemolymph was withdrawn from the pericardial sinus through a 0.2 mm hole drilled through the carapace
164 (Frederich & Pörtner, 2000). Venous blood was withdrawn from the sinus below the arthroal membrane, at the base of
165 the fourth or fifth pereopod (Greenaway & Farrelly, 1984; Giomi & Pörtner, 2013). In both cases, a small amount of
166 haemolymph (less than $20 \mu\text{l}$) was collected through capillary action using a manually sharpened Pasteur pipette with a
167 pre-inserted oxygen sensor close to the tip. Because of instrumental failure the measurement of haemolymph oxygen
168 content was only possible in Kenya.

169

170

171 *Statistical analysis*

172 For the environmental data, a Permutational Analysis of Variance was performed with the null hypothesis of no
173 differences for environmental temperatures and humidity across Regions (Kenya KY, South Africa ZA), and Sampled
174 Zones (Above ground, Below ground and Air), defined as fixed and orthogonal. Further, a PERMDISP was performed to
175 test the similarity of the variances among temperatures and humidity; whenever the variances proved heterogeneous, log-
176 transformation was applied prior to proceeding with PERMANOVA analysis. These analyses were performed using
177 PERMANOVA+ routines for PRIMER 6 (Anderson et al., 2008).

178 To test for statistical differences in MO_2 between populations and medium of respiration, an ANCOVA was performed
179 using a linear mixed model (lme4). MO_2 was set as the continuous response variable, temperature as a continuous
180 explanatory variable, and Region (Kenya, South Africa) and Medium (water, air) were set as fixed categorical explanatory
181 variables. Prior to statistical tests, MO_2 data were log transformed and the normality for each group of data was tested
182 using the Shapiro-Wilk test. Levene's test indicated homogeneity of variances in the data (d.f.=3, 69; $F=2.6$; $p=0.5883$).

183 Since measurements across the temperature ramp were made on the same individuals, they were not independent.
184 Consequently, we treated Individual ID as a random factor in the mixed model to account for multiple observations
185 (Bates, 2010).

186 The best fitted model obtained as described above, was then used to calculate the potential daily MO_2 experienced by the
187 two populations at neap and spring tide, feeding the model with the hourly average temperatures above and below ground
188 recorded in the field.

189 The same analysis was performed for haemolymph dissolved oxygen content, the only difference being that the
190 explanatory variable Region was excluded since PO_2 measurements were performed only in Kenyan and we included the
191 explanatory categorical variable Haemolymph (levels: Arterial, Venous) (Levene's Test; d.f.=3, 38; $F=3.34$; $p=0.649$).

192 The response Haemolymph oxygen content was previously square root transformed for normality. These statistical
193 analyses were carried out in R (R Development Core Team, 2014).

194

195 **Results**

196 *Tidal and thermal series*

197 The tidal range during the observation periods differed between the two study regions (Figs. 2 and 3). In Kenya, maximum
198 tidal range was 4 m during spring tides and 2.5 m during neap tides. In South Africa, the range was 2 m during spring
199 tides and 0.5 m during neap tides. Either in Kenya or South Africa, during almost all the duration of neap tides, sea level
200 did not reach the area occupied by *N. africanum* (Fig. 6).

201 Thermal regimes also differed significantly between regions: Kenya experienced less variable (PERMDISP, $t=5.67$,
202 $p<0.01$) and hotter (PERMANOVA p-hpt, $t=20.39$, $p<0.001$) temperatures than South Africa. Average temperatures in
203 Kenya ranged between 23-39°C above the sediment surface, 25-34°C below the surface and air temperatures were 22-
204 39°C. These values were significantly different from those for South Africa: 16-37°C above ground, 23-30°C below
205 ground and 15-36°C for air (PERMANOVA p-hpt, $t=7.74$, $p<0.001$ in each case)
206 Humidity measurements were not significantly different between regions (PERMDISP, $t=1.79$, $p=0.16$; PERMANOVA
207 p-hpt, $t=1.56$, $p=0.15$) with values in both localities ranging between 50% and almost 100%.

208

209 *MO₂*

210 The interaction among temperature, medium and region was highly significant (Fig. 4; $F_{4,170}=57.266$, $p<0.0001$;
211 ANCOVA), indicating that *MO₂* differed significantly between the two respiratory media (i.e. air and water) for both
212 populations. In both cases, metabolic rate was higher in air than in water, but the difference was markedly greater for
213 Kenya than South Africa. The best significant model that described the *MO₂* of *N. africanum* was represented by an
214 exponential relationship between temperature and oxygen consumption. For the Kenyan population, the best fitted model
215 during air respiration was $y = 0.018e^{0.06294k}$ with a Q_{10} of 1.8, while for respiration in water it was $0.00513e^{0.06916k}$ with a
216 Q_{10} of 1.9. For the South African population, the best models were $y = 0.0102e^{0.04358k}$ with a Q_{10} of 1.2 for air and $y =$
217 $0.000856e^{0.11070k}$ with a Q_{10} of 1.9 for water.

218

219 *Haemolymph dissolved oxygen content*

220 The oxygen content of venous and arterial haemolymph differed significantly between air and water along the temperature
221 ramp (Fig. 5; significant interaction among temperature, medium, haemolymph ANCOVA, $F_{4,139}=99.648$; $p < 0.001$). In
222 both media, *N. africanum* showed a low ability to saturate arterial haemolymph. With increasing temperatures, oxygen
223 saturation was significantly more affected in water than in air, dropping to almost 0 kPa at 33°C. During air respiration,
224 haemolymph was saturated until 34°C, dropping to 0 kPa around 37°C. Similar patterns were observed for venous
225 haemolymph, though with lower saturation levels than for arterial haemolymph.

226

227 **Discussion**

228 The thermal environment experienced by the two populations of *Neosarmatium africanum* differed significantly between
229 the two sites, with wider above ground temperature fluctuations in South Africa than in Kenya. This difference was less
230 marked for below ground temperatures, presumably because soil buffers temperature variation. Although the natural
231 thermal environment was monitored for a relatively short time, a consistent difference in temperature between latitudes

232 reflects different thermal niches across the distributional range of *N. africanum*. Above ground temperature variability
233 suggest that burrows can play a fundamental role as stable thermal refugia from the heat (Edney, 1961) that crabs of both
234 populations experience when active above ground (Sunday et al. 2014). Regardless of local differences in tidal regime,
235 both populations experience prolonged periods of emersion during neap tides, during which they are exposed to
236 fluctuations of air temperature for extended periods. Conversely, at spring tide, when the inhabited zones are flooded,
237 animals experience more stable temperature in both regions. Despite a higher variable regime of pronounced temperature
238 fluctuation, the South African population of *N. africanum* revealed a limited capability for sustaining high metabolic costs
239 under increasing temperatures. Similarly, when submerged, the crabs from the Kenyan population showed no
240 compensatory capacity for the temperature-induced increase of metabolic costs. On the contrary, while breathing in air,
241 the Kenyan crabs showed the potential to adjust their metabolism during temperature rise. It should be noted that natural
242 temperature fluctuations are more rapid and abrupt than those experimentally simulated in our protocol, and that these
243 even more severe thermal regimes may further exacerbate the different metabolic responses of the two populations. The
244 ability for such metabolic adjustment can be explained by the fact that the thermal responses are deeply influenced by the
245 biogeographic effects (Pörtner, 2001; Gaitán-Espitia et al., 2014; Baldanzi et al., 2015) that involve different thermal
246 histories for conspecific populations (Giomi et al., 2016). The tropical Kenyan population is subject to more constant, if
247 higher, temperatures than the South African population, which experiences a wider temperature range and notably low
248 temperatures during winter (Quisthoudt et al., 2013). Q_{10} values were similar for tropical specimens in air and water and
249 South African specimens in water ($Q_{10[27-37^{\circ}\text{C}]}$ = 1.8, 1.9 and 1.8 respectively), but markedly lower for South African
250 specimens in air ($Q_{10[27-37^{\circ}\text{C}]}$ = 1.2). While a Q_{10} around 2 is relatively common in marine ectotherms (Clarke & Fraser,
251 2004) and reflects a fairly normal response to increasing temperature, a Q_{10} of 1.2 indicates reduced thermal sensitivity.
252 We propose two opposing explanations for these results. The first is that, during air respiration, the animals from the
253 South African population are able to moderate its metabolic response to warming as observed in the in tropical high shore
254 snail, *Echinolittorina malaccana* (Philippi, 1847) (Marshall & McQuaid, 2011). Alternatively, we can interpret these data
255 as a sign that the South Africa population is unable to endure such acute thermal stress (Verberk et al., 2015), exhibiting
256 an inefficient metabolic response to increased temperature. This second interpretation would indicate that the temperate
257 population is more vulnerable to global warming and to heat events, in agreement with similar finding for the closely
258 related species *Perisesarma gutattum* (A. Milne Edwards, 1869) (Fusi et al., 2015).
259 The analyses of haemolymph dissolved oxygen content for the Kenyan population showed an overall decrease in oxygen
260 levels as temperature increases, with a markedly lower level of oxygen during respiration in air than water. This pattern
261 may indicate that animals from this population adjust the metabolic rate to endure increased temperatures in air
262 (Hochachka, 1991).

291 By integrating the results for temperature-dependent routine metabolic rates with the thermal data series recorded in the
292 field and the tidal regimes, we developed a diagram that predicts the daily metabolic requirements of the two study
293 populations (Fig. 6). The temperature recorded by loggers was used as body temperature of animals when above and
294 below ground during neap and spring tides and the graph shows MO_2 calculated for those temperatures. The results
295 indicate that, theoretically, Kenyan individuals are able to mobilise a wider thermal response than South African
296 conspecifics (Fig. 6). Importantly, the figure highlights the fact that, although individuals from Kenya exhibit a
297 pronounced increase of metabolic rate in air, they can rely on the cooler environment of their burrows (recorded as below
298 ground temperature), especially during spring tides, enabling them to buffer the heat load accumulated during above
299 ground activity (Edney, 1961). *N. africanum* is fully active only when its habitat within the mangrove forest is uncovered
300 by water at spring and neap tide (Micheli et al., 1991; Fratini et al., 2011). Diurnal low tides, especially during neap tides,
301 often correspond to the hottest hours of the day, maximizing the risk of thermal stress and desiccation (Porter & Gates,
302 1969; Kearney et al., 2012; Sunday et al., 2014). In contrast, the South African population has a weaker thermal response,
303 exhibiting similar oxygen consumption above and below ground.

304 We interpret the results of this study as an indication that the thermal sensitivities of the two populations differ and that
305 the temperate population in South Africa is likely to be more vulnerable to heat events, suggesting a weak capacity to
306 tolerate climate warming. In contrast, the tropical, Kenyan population shows the ability to endure heat stress by increasing
307 its respiratory rate to meet heat-induced increases in oxygen demand (Verberk et al., 2015). Indeed, tropical and tropical
308 thermal specialists may maximise their fitness within a narrow thermal niche and minimize maintenance costs, showing
309 a residual capacity for phenotypic plasticity and acclimation responses (Verberk et al., 2015). Other studies have
310 confirmed that environmental temperatures can shape thermal physiology, and that higher temperature variability
311 increases the thermal sensitivity of species (e.g. Paaijmans et al. 2013). This may be one reason why the Kenyan
312 population has evolved a more variable thermal response with a marked ability to endure higher temperatures in air than
313 the South African population. Further studies that involve other proxies such as Heat Shock Protein production,
314 behavioural assays and lactate/succinate production could confirm this (i.e. Marshall et al., 2011, 2013).

315

316 **Conclusion**

317 Our results indicate that the findings of earlier studies that tropical species are more vulnerable to climate warming than
318 temperate species is an oversimplification when considering species that span a wide range of latitudes (Deutsch et al.,
319 2008; Sunday et al., 2012) and display bimodal breathing strategy (Fusi et al., 2016). The vulnerability of species is more
320 complex and goes beyond explanations derived from general temperature envelope models based on latitudinal gradients

321 and climate, especially when species' ranges are broad so that different populations are exposed to a wide range of climatic
322 conditions. Our data show that some tropical ectotherms can show adaptation of their physiology that makes them more
323 resilient to global warming than temperate ones. This study adds data on the thermal sensitivity of intertidal tropical and
324 temperate species (Poloczanska et al. 2013) which are still largely overlooked in the estimation of the community
325 temperature index (CTI, Stuart-Smith et al. 2015), a recent and potentially powerful instrument to assess ecosystem
326 thermal vulnerability. Resilience to either heat events or chronic heating can also be highly modified by behaviour. In the
327 case of *N. africanum* this includes modulating its activity below and above ground, balancing its foraging time and burrow
328 occupancy (Sih et al., 2010; Nemeth et al., 2013) to buffer thermal stress.

329

330 **Acknowledgements**

331 The study was supported by SP3-People (Marie Curie) IRSES Project CREC (no. 247514). FG was funded by the Intra-
332 European Fellowship (ex Marie Curie) number 221017, FP7. This work is based upon research supported by the South
333 African Research Chairs Initiative of the Department of Science and Technology and the National Research Foundation.
334 MF and DD was supported also by DD baseline funding from King Abdullah University of Science and Technology
335 (KAUST). We thank Jenny Marie Booth, Sara Cilio, Bruce Mostert and Irene Ortolani for fundamental help during
336 Kenyan and South African laboratory and fieldwork.

337 **Ethical approval:**

338 All applicable international, national, and/or institutional guidelines for the care and use of animals were followed.
339 Furthermore, all procedures performed in studies involving animals were in accordance with the ethical standards of the
340 institution or practice at which the studies were conducted.

341

342 **References**

- 343 Anderson, M., R. N. Gorley, & R. K. Clarke, 2008. Permanova+ for Primer: Guide to Software and Statistical Methods.
- 344 Andretta, A., M. Fusi, I. Cameldi, F. Cimò, S. Carnicelli, & S. Cannicci, 2014. Mangrove carbon sink. Do burrowing
345 crabs contribute to sediment carbon storage? Evidence from a Kenyan mangrove system. *Journal of Sea Research* 85:
346 524–533.
- 347 Baldanzi, S., N. F. Weidberg, M. Fusi, S. Cannicci, C. D. Mcquaid, & F. Porri, 2015. Contrasting environments shape
348 thermal physiology across the spatial range of the sandhopper *Talorchestia capensis*. *Oecologia* 179: 1067–1078.
- 349 Bates, D. M., 2010. Fitting linear mixed-effects models using lme4. , <http://lme4.r-forge.r-project.org/IMMwR/lrgprt.pdf>.
- 350 Berti, R., S. Cannicci, S. Fabbroni, & G. Innocenti, 2008. Notes on the structure and the use of *Neosarmatium meinerti*
351 and *Cardisoma carnifex* burrows in a Kenyan mangrove swamp (Decapoda Brachyura). *Ethology Ecology & Evolution*
352 20: 101–113.
- 353 Bright, D. B., & C. L. Hogue, 1972. A synopsis of the burrowing land crabs of the world and list of their arthropod
354 symbionts and burrow associates. Una sinopsis mundial de los cangrejos terrestres de madrigueras y lista de sus
355 artrópodos simbios y madrigueras asociadas. *Contributions in Science* 20: 1–58.
- 356 Clarke, A., & K. P. P. Fraser, 2004. Why does metabolism scale with temperature?. *Functional Ecology* 18: 243–251.
- 357 Deutsch, C. a, J. J. Tewksbury, R. B. Huey, K. S. Sheldon, C. K. Ghalambor, D. C. Haak, & P. R. Martin, 2008. Impacts
358 of climate warming on terrestrial ectotherms across latitude. *Proceedings of the National Academy of Sciences of the*
359 *United States of America* 105: 6668–6672.
- 360 Dittmann, S., 1996. Effects of macrobenthic burrows on infaunal communities in tropical tidal flats. *Marine Ecology*
361 *Progress Series* 134: 119–130.
- 362 Duke, N., J. O. Meynecke, S. Dittmann, A. M. Ellison, K. Anger, U. Berger, S. Cannicci, K. Diele, K. C. Ewel, C. D.
363 Field, N. Koedam, S. Y. Lee, C. Marchand, I. Nordhaus, & F. Dahdouh-Guebas, 2007. A World Without Mangroves ?.
364 *Science* 317: 41–43.
- 365 Edney, E. B., 1961. The water and heat relationship of fiddler crabs (*Uca* spp.). *Transactions of the Royal Society of*

366 South Africa 36: 71–91.

367 Eliason, E. J., T. D. Clark, M. J. Hague, L. M. Hanson, Z. S. Gallagher, K. M. Jeffries, M. K. Gale, D. a Patterson, S. G.
368 Hinch, & A. P. Farrell, 2011. Differences in thermal tolerance among sockeye salmon populations. *Science* 332: 109–
369 112.

370 Fratini, S., A. Sacchi, & M. Vannini, 2011. Competition for mangrove leaf litter between two East African crabs,
371 *Neosarmatium meinerti* (Sesarmidae) and *Cardisoma carnifex* (Gecarcinidae): A case of kleptoparasitism?. *Journal of*
372 *Ethology* 29: 481–485.

373 Frederich, M., & H. O. Pörtner, 2000. Oxygen limitation of thermal tolerance defined by cardiac and ventilatory
374 performance in spider crab, *Maja squinado*. *American Journal of Physiology - Regulatory, Integrative and Comparative*
375 *Physiology* 279: R1531–R1538.

376 Fusi, M., S. Cannicci, D. Daffonchio, B. Mostert, H.-O. Pörtner, & F. Giomi, 2016. The trade-off between heat tolerance
377 and metabolic cost drives the bimodal life strategy at the air-water interface. *Scientific Reports Nature Publishing Group*
378 6: 19158.

379 Fusi, M., F. Giomi, S. Babbini, D. Daffonchio, C. D. McQuaid, F. Porri, & S. Cannicci, 2015. Thermal specialization
380 across large geographical scales predicts the resilience of mangrove crab populations to global warming. *Oikos* 124: 784–
381 795.

382 Gaitán-Espitia, J. D., L. D. Bacigalupe, T. Opitz, N. A. Lagos, T. Timmermann, & M. A. Lardies, 2014. Geographic
383 variation in thermal physiological performance of the intertidal crab *Petrolisthes violaceus* along a latitudinal gradient.
384 *The Journal of experimental biology* 217: 4379–4386.

385 Gaston, K. J., S. L. Chown, P. Calosi, J. Bernardo, D. T. Bilton, A. Clarke, S. Clusella-Trullas, C. K. Ghalambor, M.
386 Konarzewski, L. S. Peck, W. P. Porter, H. O. Pörtner, E. L. Rezende, P. M. Schulte, J. I. Spicer, J. H. Stillman, J. S.
387 Terblanche, & M. van Kleunen, 2009. Macrophysiology: a conceptual reunification. *The American naturalist* 174: 595–
388 612.

389 Gillikin, D. P., B. De Wachter, & J. F. Tack, 2004. Physiological responses of two ecologically important Kenyan
390 mangrove crabs exposed to altered salinity regimes. *Journal of Experimental Marine Biology and Ecology* 301: 93–109.

391 Gilman, E. L., J. Ellison, N. C. Duke, & C. Field, 2008. Threats to mangroves from climate change and adaptation options:
392 A review. *Aquatic Botany* 89: 237–250.

393 Giomi, F., C. Mandaglio, M. Ganmanee, G.-D. Han, Y.-W. Dong, G. A. Williams, & G. Sarà, 2016. The importance of
394 thermal history: costs and benefits of heat exposure in a tropical, rocky shore oyster. *The Journal of experimental biology*
395 219: 686–694.

396 Giomi, F., & H.-O. Pörtner, 2013. A role for haemolymph oxygen capacity in heat tolerance of eurythermal crabs.
397 *Frontiers in physiology* 4: 110.

398 Greenaway, P., & C. A. Farrelly, 1984. The venous system of the terrestrial crab *Ocypode cordimanus* (Desmarest 1825)
399 with particular reference to the vasculature of the lungs. *Journal of Morphology* 181: 133–142.

400 Hartnoll, R. G., 1975. The Grapsidae and Ocypodidae (Decapoda: Brachyura) of Tanzania. *Journal of Zoology* 177: 305–
401 328.

402 Hochachka, P. W., 1991. Temperature: the ectothermy option Phylogenetic and biochemical perspectives: *Biochemistry*

403 and molecular biology of fishes, 1. : 313–322.

404 Hoegh-Guldberg, O., & J. F. Bruno, 2010. The impact of climate change on the world's marine ecosystems. *Science* (New
405 York, N.Y.) 328: 1523–1528.

406 Huey, R. B., M. R. Kearney, A. Krockenberger, J. a M. Holtum, M. Jess, & S. E. Williams, 2012. Predicting organismal
407 vulnerability to climate warming: roles of behaviour, physiology and adaptation. *Philosophical transactions of the Royal*
408 *Society of London. Series B, Biological sciences* 367: 1665–1679.

409 Kearney, M., & W. Porter, 2009. Mechanistic niche modelling: Combining physiological and spatial data to predict
410 species' ranges. *Ecology Letters* 12: 334–350.

411 Kearney, M. R., a. Matzelle, & B. Helmuth, 2012. Biomechanics meets the ecological niche: the importance of temporal
412 data resolution. *Journal of Experimental Biology* 215: 1422–1424.

413 Kitheka, J. U., B. O. Ohowa, B. M. Mwashote, W. S. Shimbira, J. M. Mwaluma, & J. M. Kazungu, 1996. Water circulation
414 dynamics, water column nutrients and plankton productivity in a well-flushed tropical bay in Kenya. *Journal of Sea*
415 *Research.* , 257–268.

416 Kristensen, E., 2008. Mangrove crabs as ecosystem engineers; with emphasis on sediment processes. *Journal of Sea*
417 *Research* 59: 30–43.

418 Lee, S. Y., 1997. Potential trophic importance of the faecal material of the mangrove sesarmine crab *Sesarma masse*.
419 *Marine Ecology Progress Series* 159: 275–284.

420 Lee, S. Y., 2008. Mangrove macrobenthos: Assemblages, services, and linkages. *Journal of Sea Research* 59: 16–29.

421 Lovelock, C. E., D. R. Cahoon, D. A. Friess, G. R. Guntenspergen, K. W. Krauss, R. Reef, K. Rogers, M. L. Saunders,
422 F. Sidik, A. Swales, N. Saintilan, L. X. Thuyen, & T. Triet, 2015. The vulnerability of Indo-Pacific mangrove forests to
423 sea-level rise. *Nature* 526: 559-U217.

424 Macnae, W., 1968. A general account of the fauna and Flora of Mangrove Swamps and Forests in the Indo-West-Pacific
425 Region. *Advanced in marine Biology* 6: 73–270.

426 Macnae, W., & M. Kalk, 1962. The Ecology of the Mangrove Swamps at Inhaca Island, Mozambique. *Journal of Ecology*
427 50: 19–34.

428 Marshall, D. J., N. Baharuddin, & C. D. McQuaid, 2013. Behaviour moderates climate warming vulnerability in high-
429 rocky-shore snails: Interactions of habitat use, energy consumption and environmental temperature. *Marine Biology* 160:
430 2525–2530.

431 Marshall, D. J., & C. D. McQuaid, 2011. Warming reduces metabolic rate in marine snails: adaptation to fluctuating high
432 temperatures challenges the metabolic theory of ecology. *Proceedings. Biological sciences / The Royal Society* 278: 281–
433 288.

434 Marshall, D. J., C. D. McQuaid, & G. A. Williams, 2011. Non-climatic thermal adaptation: implications for species'
435 responses to climate warming. *Biology letters* 7: 160.

436 Micheli, F., F. Gherardi, & M. Vannini, 1991. Feeding and burrowing ecology of two East African mangrove crabs.
437 *Marine Biology* 111: 247–254.

438 Nemeth, Z., F. Bonier, & S. MacDougall-Shackleton, 2013. Coping with Uncertainty: Integrating Physiology, Behavior,

439 and Evolutionary Ecology in a Changing World. *Integrative and Comparative Biology* 53: 960–964.

440 Paaïjms, K. P., R. L. Heinig, R. a. Seliga, J. I. Blanford, S. Blanford, C. C. Murdock, & M. B. Thomas, 2013.
441 Temperature variation makes ectotherms more sensitive to climate change. *Global Change Biology* 19: 2373–2380.

442 Poloczanska, E. S., C. J. Brown, W. J. Sydeman, W. Kiessling, D. S. Schoeman, P. J. Moore, K. Brander, J. F. Bruno, L.
443 B. Buckley, M. T. Burrows, C. M. Duarte, B. S. Halpern, J. Holding, C. V. Kappel, M. I. O’Connor, J. M. Pandolfi, C.
444 Parmesan, F. Schwing, S. A. Thompson, & A. J. Richardson, 2013. Global imprint of climate change on marine life.
445 *Nature Climate Change* 3: 919–925.

446 Porter, W., & D. Gates, 1969. Thermodynamic Equilibria of Animals with Environment. *Ecological monographs* 39:
447 227–244.

448 Pörtner, H., 2001. Climate change and temperature-dependent biogeography: Oxygen limitation of thermal tolerance in
449 animals. *Naturwissenschaften* 88: 137–146.

450 Quisthoudt, K., J. Adams, A. Rajkaran, F. Dahdouh-Guebas, N. Koedam, & C. F. Randin, 2013. Disentangling the effects
451 of global climate and regional land-use change on the current and future distribution of mangroves in South Africa.
452 *Biodiversity and Conservation* 22: 1369–1390.

453 R Development Core Team, 2014. R: A Language and Environment for Statistical Computing. <http://www.R-project.org>.
454 R Foundation for Statistical Computing, Vienna, <http://www.r-project.org>.

455 Ragionieri, L., S. Cannicci, C. D. Schubart, & S. Fratini, 2010. Gene flow and demographic history of the mangrove crab
456 *Neosarmatium meinerti*: A case study from the western Indian Ocean. *Estuarine, Coastal and Shelf Science* 86: 179–188.

457 Ragionieri, L., S. Fratini, & C. D. Schubart, 2012. Revision of the *Neosarmatium meinerti* species complex (Decapoda:
458 Brachyura: Sesarmidae), with descriptions of three pseudocryptic Indo–West Pacific species. *The Raffles Bulletin of*
459 *Zoology* 60: 71–87.

460 Rajkaran, A., & J. Adams, 2012. The effects of environmental variables on mortality and growth of mangroves at
461 Mngazana Estuary, Eastern Cape, South Africa. *Wetlands Ecology and Management* 20: 297–312.

462 Rezende, E. L., L. E. Castaneda, & M. Santos, 2014. Tolerance landscapes in thermal ecology. *Functional Ecology* 28:
463 799–809.

464 Ridd, P. V., 1996. Flow Through Animal Burrows in Mangrove Creeks. *Estuarine, Coastal and Shelf Science* 43: 617–
465 625.

466 Roznik, E. a., & R. a. Alford, 2012. Does waterproofing Thermochron iButton dataloggers influence temperature
467 readings?. *Journal of Thermal Biology Elsevier* 37: 260–264, <http://dx.doi.org/10.1016/j.jtherbio.2012.02.004>.

468 Schurmann, H., & J. F. Steffensen, 1992. Lethal oxygen levels at different temperatures and the preferred temperature
469 during hypoxia of the Atlantic cod, *Gadus morhua* L. *Journal of Fish Biology*. , 927–934.

470 Sih, A., J. Stamps, L. H. Yang, R. McElreath, & M. Ramenofsky, 2010. Behavior as a key component of integrative
471 biology in a human-altered world. *Integrative and Comparative Biology* 50: 934–944.

472 Sinclair, B. J., K. E. Marshall, M. A. Sewell, D. L. Levesque, C. S. Willett, S. Slotsbo, Y. Dong, C. D. G. Harley, D. J.
473 Marshall, B. S. Helmuth, & R. B. Huey, 2016. Can we predict ectotherm responses to climate change using thermal
474 performance curves and body temperatures?. *Ecology Letters* 19: 1372–1385.

475 Smith, T. J., K. G. Boto, S. D. Frusher, & R. L. Giddins, 1991. Keystone species and mangrove forest dynamics: the
476 influence of burrowing by crabs on soil nutrient status and forest productivity. *Estuarine, Coastal and Shelf Science* 33:
477 419–432.

478 Stieglitz, T., P. Ridd, & P. Müller, 2000. Passive irrigation and functional morphology of crustacean burrows in a tropical
479 mangrove swamp. *Hydrobiologia* 421: 69–76.

480 Stuart-smith, R. D., G. J. Edgar, N. S. Barrett, S. J. Kininmonth, & A. E. Bates, 2015. Thermal biases and vulnerability
481 to warming in the world ' s marine fauna. *Nature* Nature Publishing Group 1–17.

482 Sunday, J. M., A. E. Bates, & N. K. Dulvy, 2011. Global analysis of thermal tolerance and latitude in ectotherms.
483 *Proceedings. Biological sciences / The Royal Society* 278: 1823–1830.

484 Sunday, J. M., A. E. Bates, & N. K. Dulvy, 2012. Thermal tolerance and the global redistribution of animals. *Nature*
485 *Climate Change* Nature Publishing Group 2: 686–690.

486 Sunday, J. M., A. E. Bates, M. R. Kearney, R. K. Colwell, N. K. Dulvy, J. T. Longino, & R. B. Huey, 2014. Thermal-
487 safety margins and the necessity of thermoregulatory behavior across latitude and elevation. *Proceedings of the National*
488 *Academy of Sciences of the United States of America* 111: 5610–5615.

489 Tack, J., D. P. Gillikin, & S. De Grave, 2001. The occurrence of the semi-terrestrial shrimp *Merguia oligodon* (De Man,
490 1888) in *Neosarmatium smithi* (H. Milne Edwards, 1853) Burrows in kenyan mangroves. *Crustaceana* 74: 505–507.

491 Terblanche, J. S., A. A. Hoffmann, K. A. Mitchell, L. Rako, P. C. Roux, & S. L. Chown, 2011. Ecologically relevant
492 measures of tolerance to potentially lethal temperatures. 10: 3713–3725.

493 Tomlinson, P. B., 1986. The botany of mangroves. Cambridge tropical biology series.

494 Verberk, W. C. E. P., F. Bartolini, D. J. Marshall, H.-O. Pörtner, J. S. Terblanche, C. R. White, & F. Giomi, 2015. Can
495 respiratory physiology predict thermal niches?. *Annals of the New York Academy of Sciences* 1–16.

496 Verberk, W. C. E. P., J. Overgaard, R. Ern, M. Bayley, T. Wang, L. Boardman, & J. S. Terblanche, 2016. Does oxygen
497 limit thermal tolerance in arthropods? A critical review of current evidence. *Comparative Biochemistry and Physiology -*
498 *Part A : Molecular and Integrative Physiology* The Authors 192: 64–78.

499 Warren, J. H., & a. J. Underwood, 1986. Effects of burrowing crabs on the topography of mangrove swamps in New
500 South Wales. *Journal of Experimental Marine Biology and Ecology* 102: 223–235.

501 Williams, S. E., L. P. Shoo, J. L. Isaac, A. a Hoffmann, & G. Langham, 2008. Towards an Integrated Framework for
502 Assessing the Vulnerability of Species to Climate Change. *PLoS biology* 6: 6.

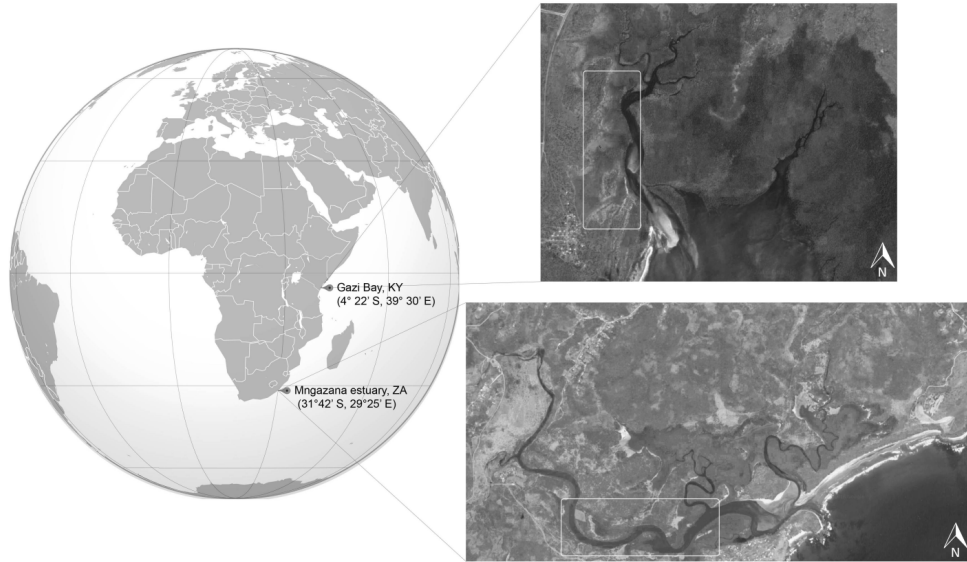
503

504

505 **Figures**

506

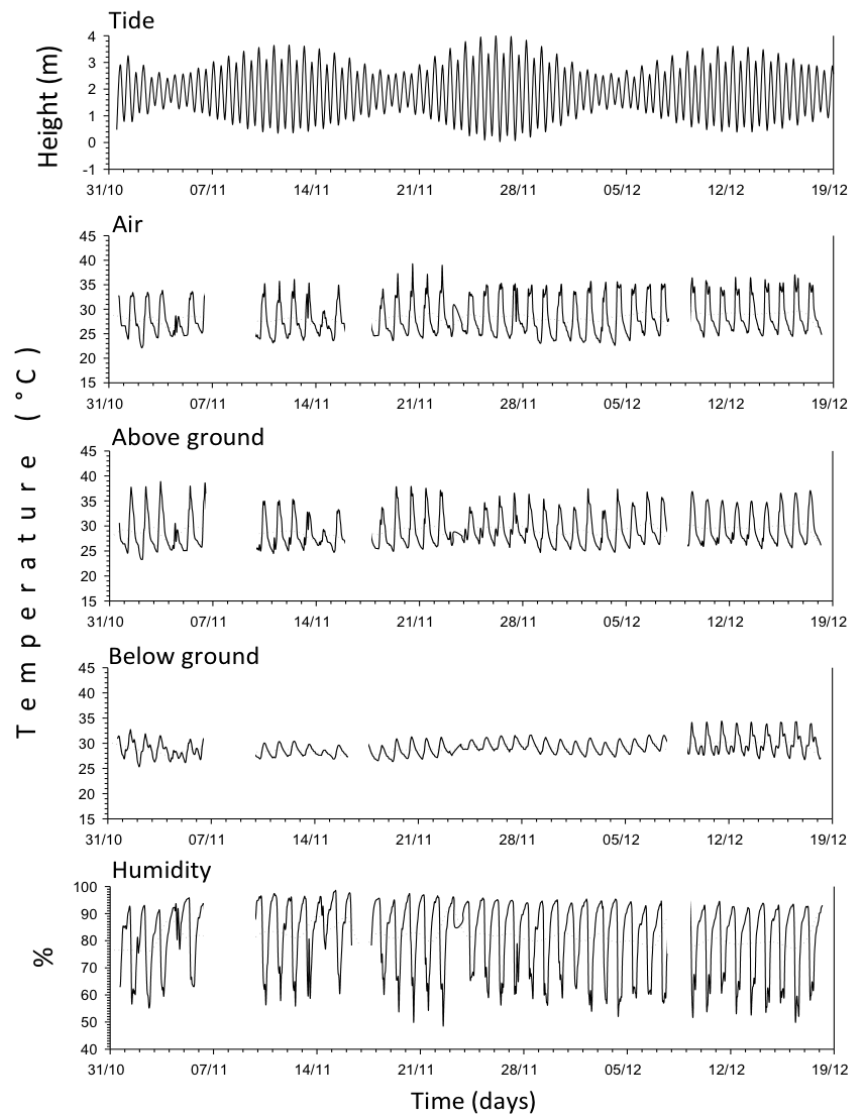
507 **Fig. 1.** Study sites along the east coast of Africa. The tropical one at Gazi, Kenya, and temperate one, Mngazana, South
508 Africa.



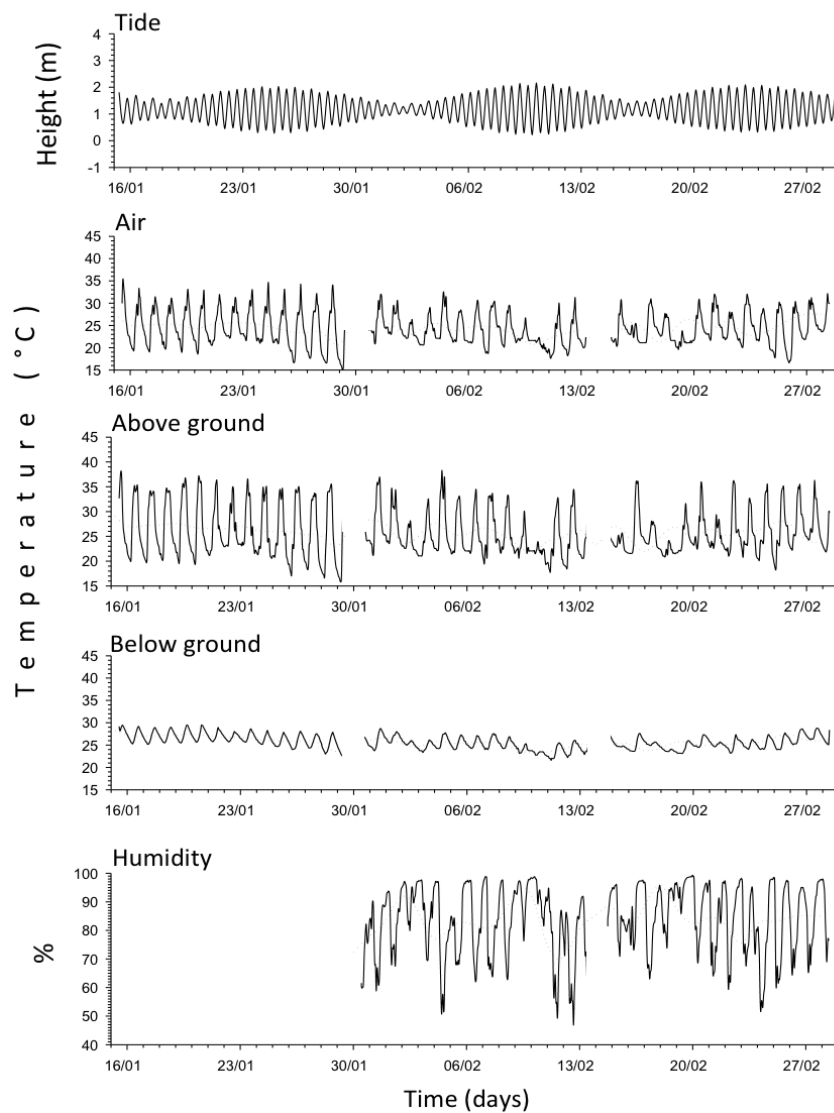
509

510

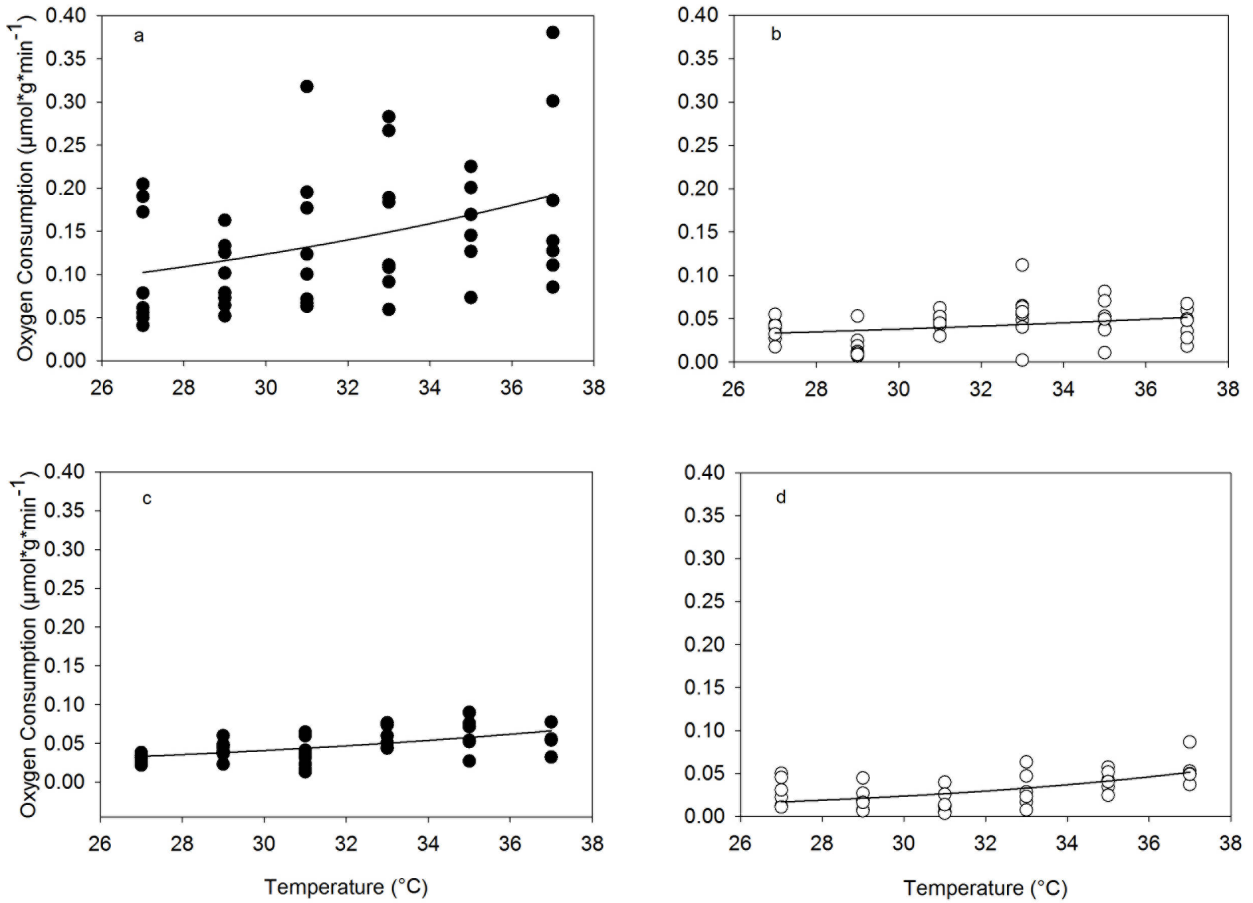
511 **Fig. 2.** Environmental temperatures and humidity of the area colonized by *Neosarmatium africanum* in Kenya during the
512 period 31 October to 19 December 2011. The dashed line is the daily average of the variable described in each graph,
513 while the solid line is the hourly average.



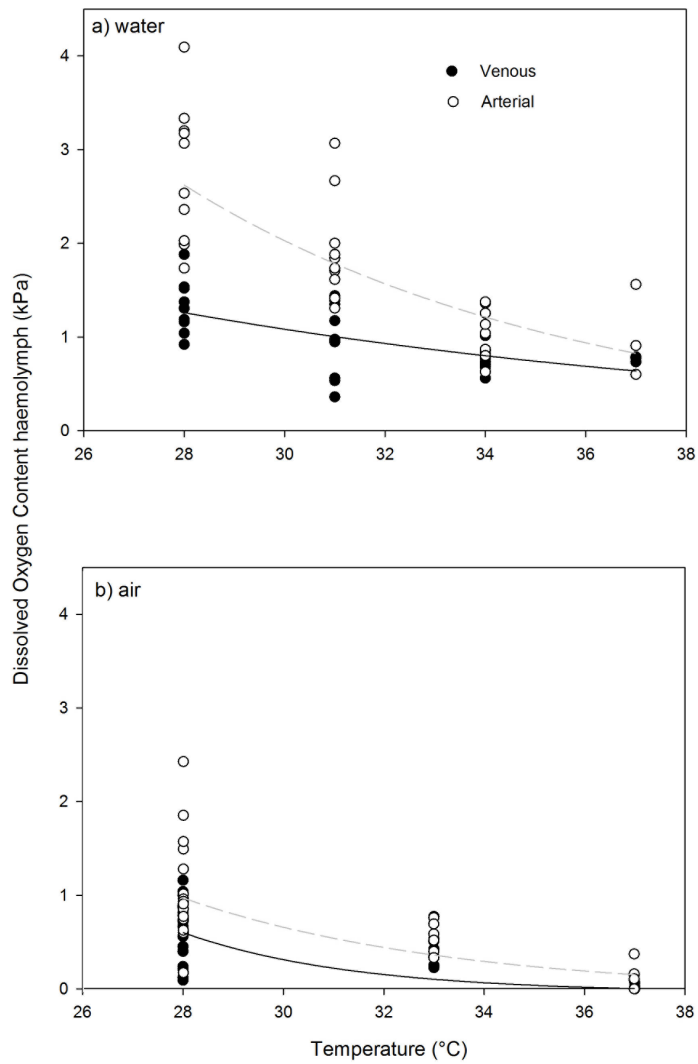
515 **Fig. 3.** Environmental temperature and humidity of the area colonized by *Neosarmatium africanum* in South Africa during
516 the period 16 January to 5 March 2012. The dashed line is the daily average of the variable described in each graph, while
517 the solid line is the hourly average.



519 **Fig. 4.** MO_2 of *Neosarmatium africanum* from Kenya in air (a) and water (c) and from South Africa in air (b) and water
520 (d). The significant best fitted models (see results section for the equations) are represented with continuous black lines
521 for each population and medium.



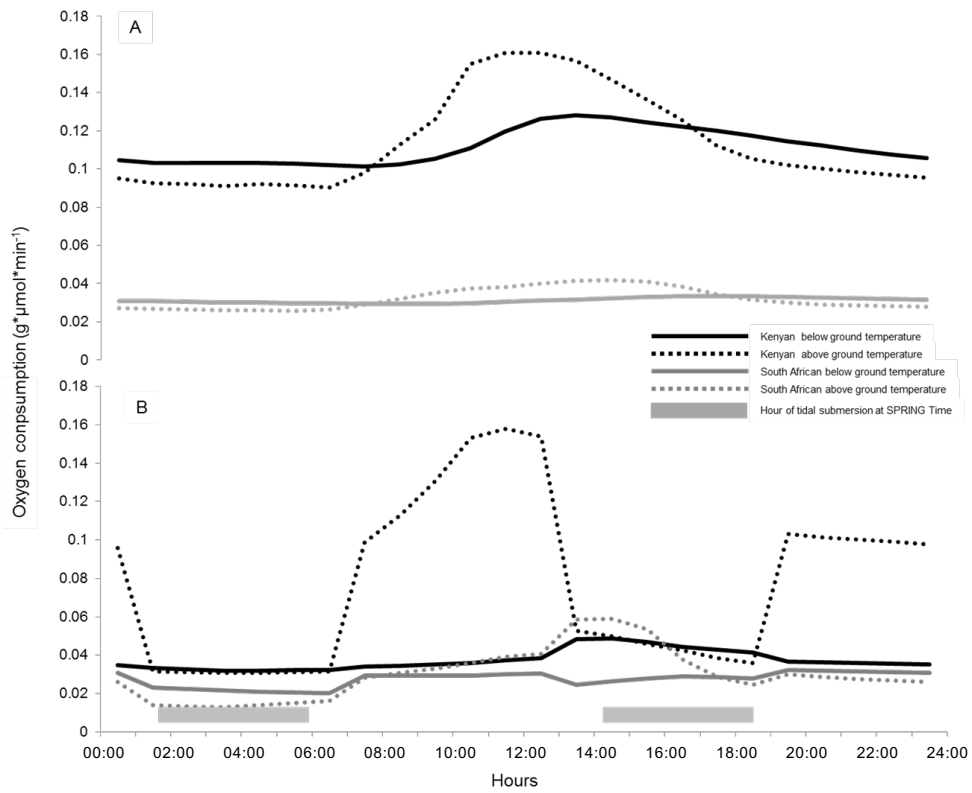
523 **Fig. 5.** Haemolymph dissolved oxygen content of Kenyan *Neosarmatium africanum* during water (A) and air (B)
 524 respiration; open circles arterial haemolymph, black circles venous haemolymph. Significant regressions are plotted:
 525 arterial ($y = 96.016e^{-0.129κ}$ for water and $176.49e^{-0.0185κ}$ for air respiration) shown in dotted-grey line, while venous ($y =$
 526 $10.516e^{-0.076κ}$ for water and $6058.2e^{-0.0332κ}$ for air respiration) shown in solid-black line.



527

528

529 **Fig. 6.** Hourly MO_2 predicted for *Neosarmatium africanum* during Neap Tide (A) and Spring Tide (B) on the basis of
 530 temperatures measured above ground and below ground (see the legend in the graph). Shaded bars indicate the hours
 531 flooded at both sites during spring tides.



532

533

534